

**CALIFORNIA DEPARTMENT OF FOOD AND AGRICULTURE
FERTILIZER RESEARCH AND EDUCATION PROGRAM (FREP)**

FINAL REPORT

Project Title

Minimizing nitrogen runoff and improving nitrogen use efficiency in containerized woody ornamentals through management of nitrate and ammonium-nitrogen

Project Number

00-0509

Project Location

UC Davis and UC Riverside

Project Duration

2.6 years

Project Leaders

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A. PROJECT SUMMARY

The main objectives of this research were to develop fertilization and irrigation guidelines for woody ornamental crop production that will minimize nitrate (NO₃) runoff and improve nitrogen (N) use efficiency (NUE). Two major experiments were done to characterize the dynamics of N cycling in the plants and the media.

University of California – Davis. A hydroponics study was conducted at Davis, California. By monitoring the rates of N and water depletion from nutrient solutions, the researchers were able to characterize the dynamics of N and water demand for several ornamental crops, as affected by physiological (stage of plant development) and environmental conditions (time of year).

University of California - Riverside. The study conducted at Riverside, California investigated the fate of different controlled-release fertilizers (CRF) and liquid fertilizers (LF) as affected by acid pH (5.0) media in a temperature-controlled

greenhouse setting and neutral pH (7.0) media in an outdoor setting (no temperature control). The dynamics of N cycling in the planting media and N uptake into the plants were determined by measuring nitrate (NO_3^-) and ammonium (NH_4^+) leaching from containers on a weekly basis and extractable NH_4^+ , NO_3^- and total N in the planting media and total N in plants on a monthly basis. Plant response to fertilizers was determined by measuring total N accumulation in the plants on a monthly basis.

Specific Objectives:

1. Characterize the N and water demands of container-grown ornamental plants as influenced by plant growth rate, stage of plant development, and environmental conditions.
2. Determine the fate of NH_4^+ and NO_3^- from CRF and LF in containerized woody ornamentals growing in acid (5.0) or neutral (7.0) pH media during a 12-month period.
3. Develop fertilization and irrigation guidelines based on research results. Actively distribute guidelines to growers, CE advisors, consultants, fertilizer companies and educators through workshops, field days, seminars, lectures, and publications.

Added Objectives

In addition to the two well-known pollutants, nitrate (NO_3^-) and phosphate (PO_4^{2-}), other essential plant nutrients listed in §101(a) of the Clean Water Act (U.S.

Environmental Protection Agency, 1994) are boron (B), copper (Cu), iron (Fe), manganese (Mn) and zinc (Zn). Although not receiving as much attention as NO_3^- and PO_4^{2-} , these plant micronutrients are chemicals that are used by the nursery industry and, therefore, are at risk of being regulated by state agencies.

The likelihood of micronutrient contamination of surface waters from nursery runoff is unknown. Therefore, information regarding the movement of these nutrients in nursery production systems is needed so that the nursery industry is aware of any potential problems associated with micronutrient fertilization so that

Best Management Practices (BMPs) may be developed to mitigate the runoff of these nutrients. Since most essential plant nutrients are discussed as potential pollutants in EPA guidelines, and we were collecting samples of leachates, media and plant tissue, we decided to monitor all essential plant nutrients, with the exception of sulfur.

The results of the research funded by this grant are presented in two sections: (1) the first task presents the data and results regarding nitrogen and water use, which were conducted by Dr. Richard Evans at Davis, California and (2) the second task presents the data and results regarding the use of Controlled Release Fertilizers (CRF) and liquid fertilization programs.

Task 1

Water and Nitrogen Demand of Container-Grown Woody Plants

Month of initiation: 1/2001

Month of completion: 2/2002

Subtask 1.1: A static solution culture system was prepared in an outdoor nursery at the Environmental Horticulture Department at Davis, CA (Figure 1). An aeration system, consisting of a main manifold with individual air lines directed into each container, was constructed. Completed by 2/2001.



Figure 1. Static solution culture system at U.C. Davis.

Subtask 1.2: A modified, half-strength Hoagland's solution with full strength minor nutrients was prepared for the system. N form was adjusted as necessary to control pH. Completed by 2/2001.

Subtask 1.3: Five woody ornamental species (*Berberis thunbergii*, *Hydrangea macrophylla*, *Ilex aquifolium*, *Ligustrum lucidum* and *Rhododendron* sp.) were

obtained as 2-inch liners. Roots were carefully cleaned of media and placed in nutrient solutions. Completed by 3/2001.

Subtask 1.4: Nitrogen and water uptake were monitored. Nutrient solutions were changed at three-day intervals, or as needed. Water use was determined gravimetrically on each container. N uptake was calculated by measuring N depletion from nutrient solutions each time solutions were changed. Nitrogen concentration of the solutions was determined by the diffusion-conductivity method (Carlson et al., 1990). Plant fresh weight and shoot elongation were measured weekly. Subtask was completed by 9/2001.

TASK 1 - RESULTS AND CONCLUSIONS

Five woody ornamental species (*Berberis thunbergii*, *Hydrangea macrophylla*, *Ilex aquifolium*, *Ligustrum lucidum*, and *Rhododendron sp.*) were obtained in Spring 2001 as 2-inch liners and planted into 4-L static solution culture containers and placed on benches in a lath-house at the Department of Environmental Horticulture. The composition of the nutrient solution used during most of the experimental period was 2 mM MgSO₄, 1 mM K₂SO₄, 1 mM KH₂PO₄, 2 mM NH₄NO₃, CaSO₄ at 0.43 g/L, and micronutrients at full-strength Hoagland's solution concentrations. Acidity was adjusted to pH 6, and solutions were changed every 3-4 days to maintain sufficient amounts of water and nutrients in the containers. From day 128 to day 138, the solution for half of the plants of each species was maintained between pH 4.5-5. The weight of the nutrient solution in each container was determined before and after each solution change for calculation of water use and N uptake. The concentration of NO₃-N and NH₄-N in the nutrient solution was determined before and after each change by the diffusion-conductivity method. Plant fresh weight was determined at each nutrient solution change after lightly blotting the roots to remove excess nutrient solution water.

Cumulative total N uptake varied greatly by species (Fig. 1). Total N uptake by *Hydrangea* was twice as great as uptake by any other species. This difference was even more pronounced for NO₃-N uptake (Fig. 2). With the exception of *Hydrangea*, all of the species under study took up more NH₄-N than NO₃-N (Fig. 3). Lowering the solution pH did not significantly affect plant preference for NH₄-N or NO₃-N.

During the first 50 days of growth, average daily water uptake of all five species was 50-75 mL per day (Fig. 4). After about 3 months of growth, average water uptake rates for most of the species were between about 160-225 mL per day (Table 1). The exception was *Hydrangea*, for which the average had increased to 470 mL per day. These rates of water use are lower than estimated irrigation application rates at most commercial nurseries.

The ratio of N uptake to water uptake yields a value for the ideal nutrient solution N concentration for a liquid feed system (Table 1). The highest ratio occurred in *Hydrangea* (54 mg/L) and the lowest in *Rhododendron* (23 mg/L). All

of these values are substantially lower than the liquid feed N concentrations applied in most commercial nurseries.

The growth habits and soil preferences of the species studied are representative of the range typically found in commercial nurseries. The results will be useful to growers who seek finer adjustment of application rates of nitrogen and water, as well as to fertilizer companies and others who wish to match nitrogen application rates or release rates to woody ornamental crop needs.

Table 1. Average daily uptake of N (in mg), water (in mL), and the ratio of total N uptake to water uptake, in mg/L.

	<u>mg N</u>	<u>mL water</u>	<u>mg N/L</u>
Berberis	6.4	160	39.9
Hydrangea	25.3	470	54.0
Ilex	5.1	169	30.2
Ligustrum	11.0	226	48.7
Rhododendron	4.4	187	23.3

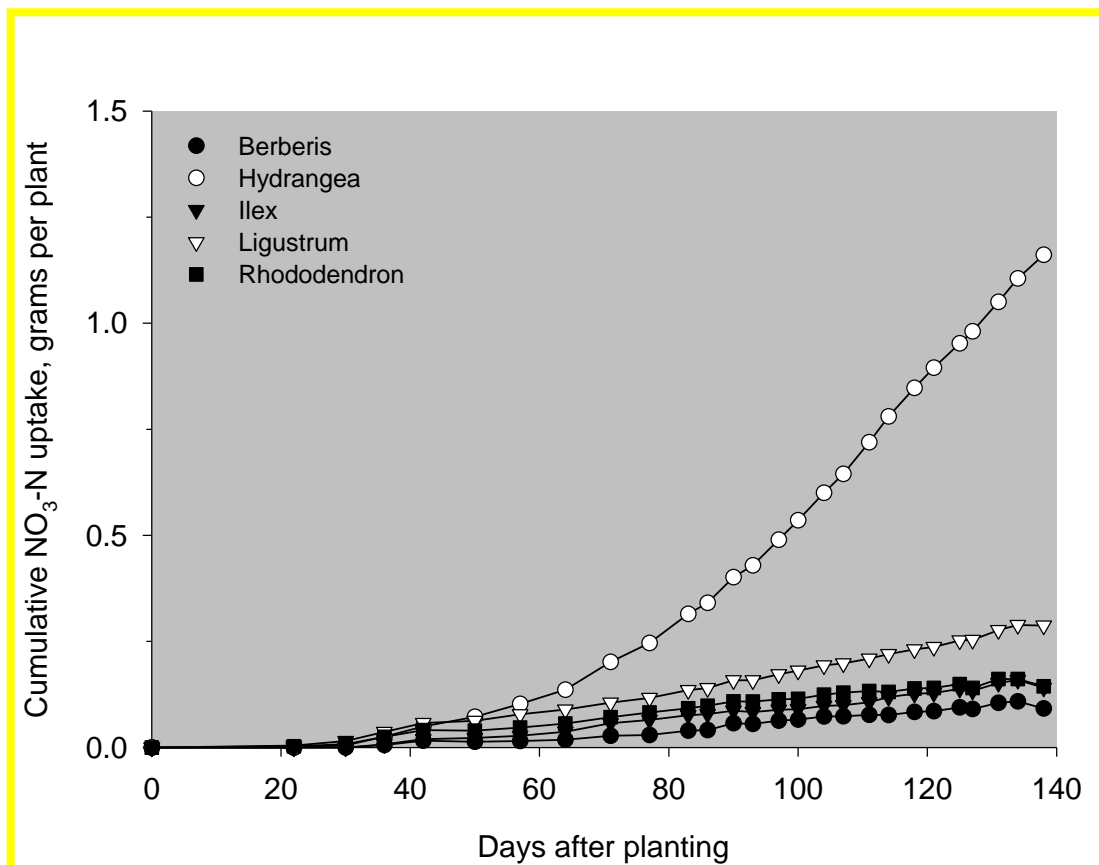


Figure 1. Cumulative nitrogen (N) uptake for *Berberis*, *Hydrangea*, *Ilex*, *Ligustrum*, and *Rhododendron*.

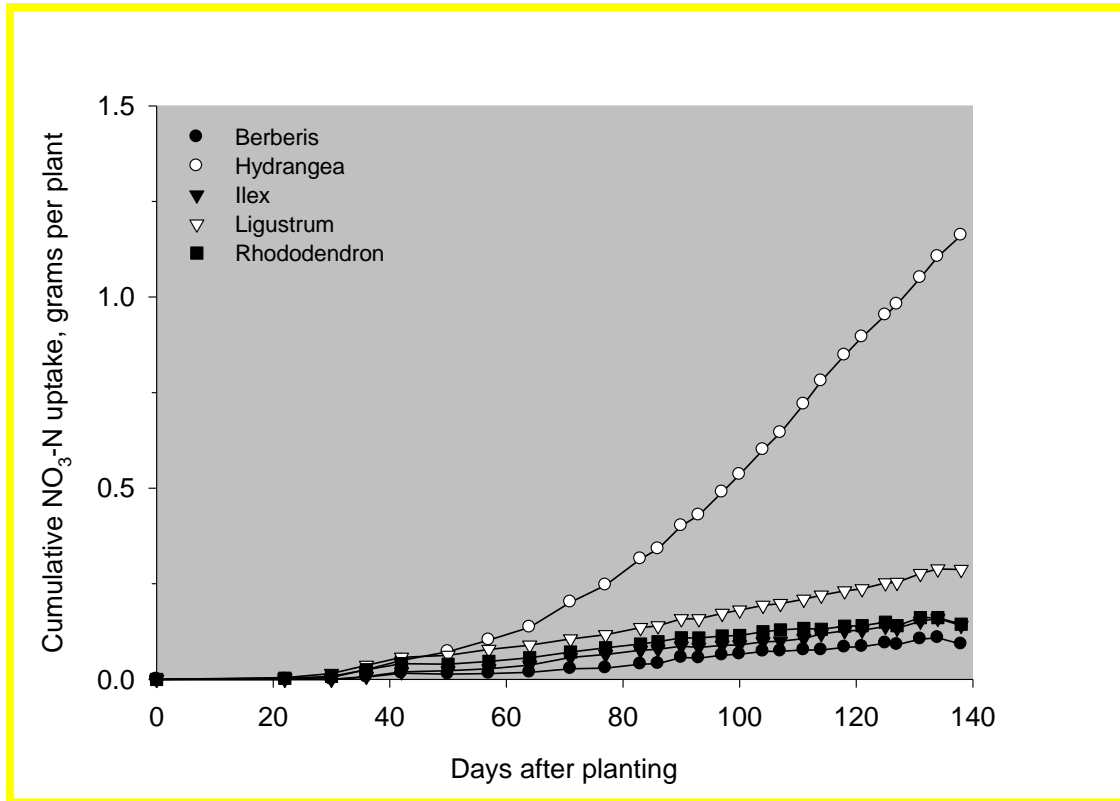


Figure 2. Cumulative $\text{NO}_3\text{-N}$ uptake of *Berberis*, *Hydrangea*, *Ilex*, *Ligustrum*, and *Rhododendron*.

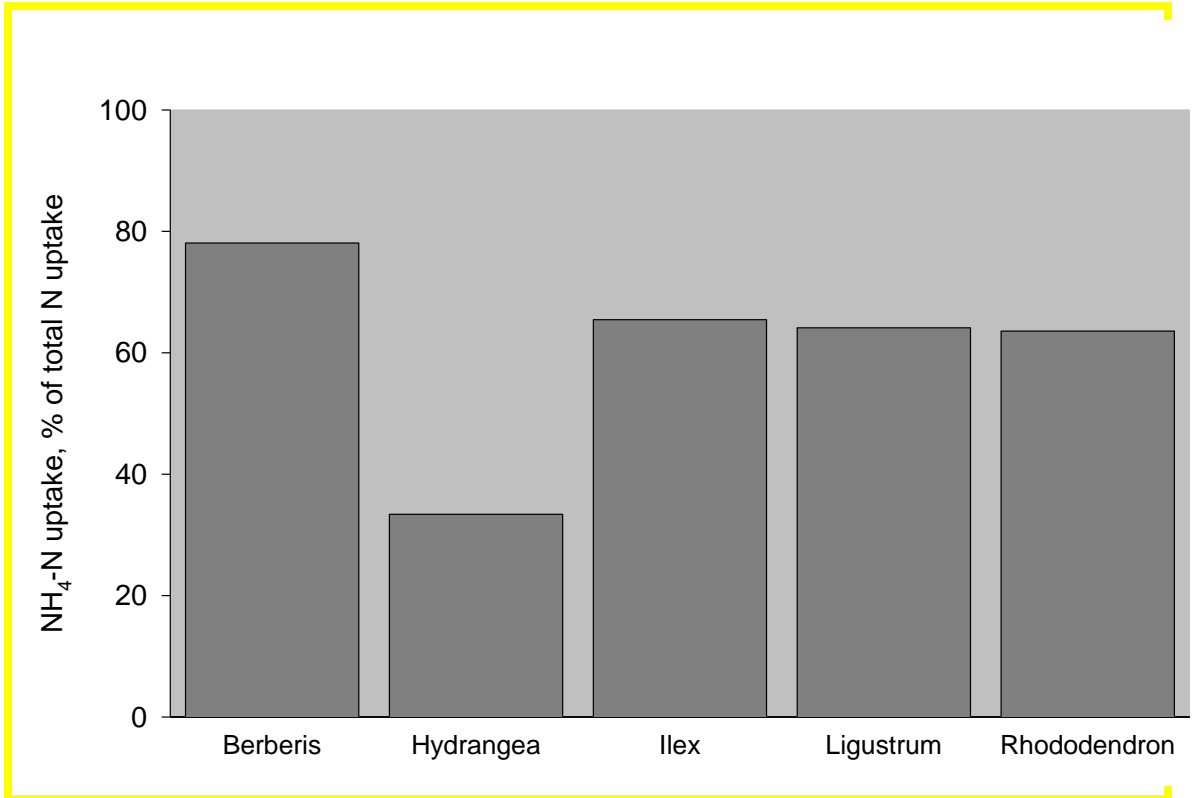


Figure 3. Ammonium (NH₄⁺-N) uptake as a percentage of total N uptake in *Berberis*, *Hydrangea*, *Ilex*, *Ligustrum* and *Rhododendron*.

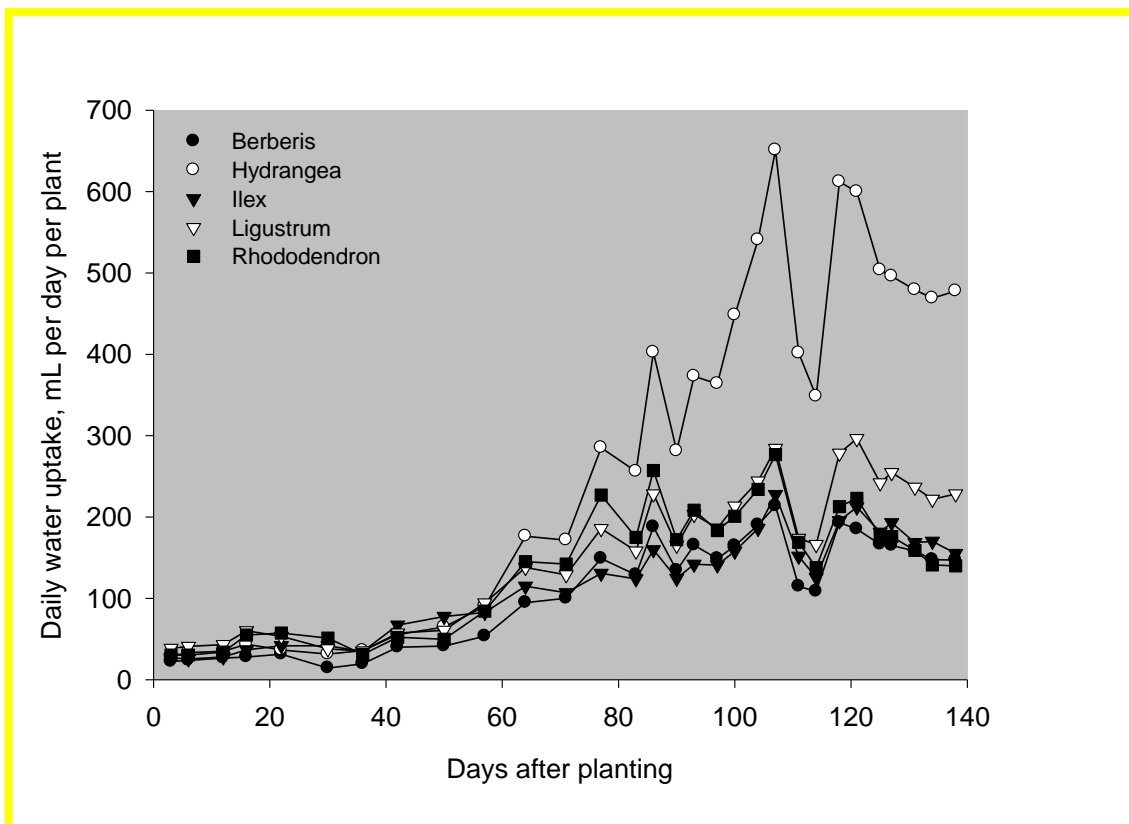


Figure 4. Daily water uptake rates in *Berberis*, *Hydrangea*, *Ilex*, *Ligustrum* and *Rhododendron*

Task 2

Ammonium, Nitrate, Phosphorus, Potassium, Calcium, Magnesium, Iron, Manganese, Copper, Zinc, and Molybdenum Release Characteristics from Containerized Acid Media in a Temperature-Regulated Greenhouse

Task 2. Determine the fate of NH_4^+ and NO_3^- from CRF and LF in containerized woody ornamentals growing in acid (5.0) or neutral (7.0) pH media throughout an 11-month period.

Month of Initiation: 1/2001

Month of completion: 7/2002

Subtask 2.1: Experimental plots were laid out and prepared for the installation of benches and the irrigation system. Treatments were set up into 10 blocks for the ligustrum crop (5 blocks for the plant study and 5 blocks for the controls – no plants) and 10 blocks for the azalea crop (5 blocks for the plant study and 5 blocks for the controls – no plants). Each of the 7 treatments was randomly assigned a position in each block (Figure 5). There were 11 replications of each treatment in each block, with one replication being harvested each month. Shade cloth was purchased. All fertilizer, irrigation equipment, and bench materials were purchased.

Completed by 2/2001

Osmo/7	Nutri/7	NO3/9	NH4NO3/6	NH4/9	NH4NO3/11	NH4/10	NO3/5
B10L1	B10L2	B10L3	B10L4	B10L5	B10L6	B10L7	B10L8
Apex/2	Nutri/9	Nutri/4	Multi/7	NH4/7	Apex/10	NH4NO3/3	NO3/10
B10L9	B10L10	B10L11	B10L12	B10L13	B10L14	B10L15	B10L16
Multi/1	NH4NO3/9	Apex/11	Nutri/3	NH4NO3/7	NH4NO3/2	Multi/8	NH4/1
B10L17	B10L18	B10L19	B10L20	B10L21	B10L22	B10L23	B10L24
Nutri/10	NH4/8	Multi/2	NO3/8	Osmo/3	NH4NO3/5	Nutri/11	NO3/4
B10L25	B10L26	B10L27	B10L28	B10L29	B10L30	B10L31	B10L32
NO3/11	NH4NO3/10	Osmo/8	Multi/3	Nutri/8	Apex/6	Osmo/11	NO3/1
B10L33	B10L34	B10L35	B10L36	B10L37	B10L38	B10L39	B10L40
NO3/6	NH4/5	Osmo/6	Apex/3	Osmo/2	Osmo/4	Nutri/2	Apex/4
B10L41	B10L42	B10L43	B10L44	B10L45	B10L46	B10L47	B10L48
NH4NO3/1	Osmo/1	Osmo/9	NH4/2	Multi/6	Multi/4	Multi/11	Nutri/6
B10L49	B10L50	B10L51	B10L52	B10L53	B10L54	B10L55	B10L56
Apex/5	Apex/7	Multi/5	NO3/7	NO3/3	Apex/1	Apex/9	NH4NO3/4
B10L57	B10L58	B10L59	B10L60	B10L61	B10L62	B10L63	B10L64
Nutri/1	Nutri/5	NH4NO3/8	Multi/10	NH4/3	NH4/6	NO3/2	Apex/8
B10L65	B10L66	B10L67	B10L68	B10L69	B10L70	B10L71	B10L72
Multi/9	NH4/11	Osmo/10	NH4/4	Osmo/5			
B10L73	B10L74	B10L75	B10L76	B10L77	B10L78	B10L79	B10L80

Figure 5. A sample of a block, showing all of the treatments that were randomly assigned a location within the block. The layout description consisted of the treatment name: Osmocote (Osmo), Nutricote (Nutri), Multicote (Multi), Apex, liquid-fertilizer nitrate (NO_3^-), liquid-fertilizer ammonium (NH_4^+) and liquid-fertilizer ammonium nitrate ($\text{NH}_4^+\text{NO}_3^-$). Month of harvest is marked to the right of the treatment name (/1, /2, etc.). The letter and number code below the treatment name specifies the block and bench location. For example, B10L73 indicates the plant is at block 10 and location 73 on the bench.

Subtask 2.2: Benches, irrigation, and drainage systems were built and installed (Figures 6 and 7). The irrigation system was designed for automated irrigation of all treatments. There were four different irrigation lines (Figure 8) – city water was used for irrigating all CRF treatments, and three separate water lines represented the three different liquid fertilizer treatments (Figure 9). Materials needed to collect leachate were ordered and adapted to the system to collect leachate.

Completed by 3/2001.



Figure 6. Greenhouse benches that were used for azalea production. Photo shows Blocks 1-5, with Block 1 in the foreground and Block 5 in the background. Shade cloth (33% shade) covers the structure during the months of May to October.



Figure 8. Four irrigation lines used to irrigate crops. Drippers extended out from each line to the pot to be irrigated. White-taped lines were water for the controlled-release fertilizer treatments and the orange, pink and yellow-taped lines represented the three liquid fertilizer treatments.



Figure 7. Greenhouse benches that were used for ligustrum production. Photo shows Blocks 6-10, with Block 6 in the foreground and Block 10 in the background. Plants were not provided any protection from weather.



Figure 9. Three Dosatron pumps used to portion out fertilizer for the three liquid fertilizer treatments.

Subtask 2.3: A total of 840 plants (*Azalea Southern Indica* ‘Phoenicia’, and *Ligustrum texanum*) at the liner-stage (Figure 10), were obtained from a commercial nursery. Plants were potted into 1-gallon containers containing the

appropriate media (Figure 11). In the case of the Controlled Release Fertilizers (CRF), fertilizers were thoroughly mixed into the media with a cement mixer. Treatments were a 2 x 7 factorial of 2 different media pH (5.0 and 7.0) and seven different fertilizer treatments. The liquid-fertilizer study was initiated in March 2002. There were five replications of each treatment for each of the 11 monthly harvests. Substrates for the media were purchased and delivered. Completed by 4/2001.



Figure 10. Liners of plants prior to being planted into 1-gallon sized plastic containers.



Figure 11. Liners being planted into containers. All controlled-release fertilizers were mixed into media with a cement mixer.

1 List of fertilizer treatments for Task 2.
2

<u>Treatment</u>	<u>Fertilizer</u>
3	
4	
5	
6 1	100 ppm N as 75% NH ₄ ⁺ and 25% NO ₃ ⁻ Liquid fertilizer
7 2	100 ppm N as NH ₄ NO ₃ Liquid Feed
8 3	100 ppm N as 75% NO ₃ and 25% NH ₄ ⁺ Liquid fertilizer
9 4	Osmocote CRF
10 5	Apex CRF
11 6	Multicote CRF
12 7	Nutricote CRF

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17 **Subtask 2.4:** Baseline measurements of media (without N fertilizer) and plants were
18 taken. Five replications each of the acid and neutral pH media were measured for total
19 KCL-extractable NH₄⁺ and NO₃⁻, total N, electrical conductivity (EC) pH and total nutrient
20 content. Concentrations of NH₄⁺ and NO₃⁻ were determined with a Technicon Continuous
21 Flow Analyzer. Five each of the Azalea Southern Indica 'Phoenicia' and the *Ligustrum*
22 *texanum* were harvested, separated into roots and shoots, dried and ground. Dry weights
23 and N concentration of roots and shoots were measured. This data has been collected.
24 This subtask for the controlled release fertilizers was completed by the end of 7/2001. The
25 Liquid Fertilizer portion was completed by the end of 4/2002.

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27 **Subtask 2.5 and 2.6:** Weekly tasks of leachate collection and analyses (EC,
28 NO₃⁻-N content and NH₄⁺-N content) were performed as planned.
29

30 **TASK 2 – CURRENT RESULTS AND DISCUSSION**

31 All experiments are completed. However, data is still being processed. Therefore,
32 the results of the Controlled Release Fertilizer (CRF) studies and the Liquid
33 Fertilization (LF) studies will be presented in 6 parts:
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35
36 1. Release patterns of plant nutrients from four controlled release fertilizers in
37 **Acid** media during an eleven-month period in a simulated greenhouse production
38 facility – A control study (no plants).
39

40 *The results of this study are presented in two manuscripts at the end of this*
41 *document.*
42

43 2. Release patterns of plant nutrients from four controlled release fertilizers in
44 **Neutral** media during an eleven month period in a simulated outdoor production
45 facility – A control study (no plants)

- 46 3. Release of plant nutrients from four **controlled release fertilizers** and
47 leaching and plant uptake characteristics for **Greenhouse-Grown Azaleas**
48 during an eleven month period.
49
- 50 4. Release of plant nutrients from four **controlled release fertilizers** and
51 leaching and plant uptake characteristics for **Outdoor-Grown Ligustrum** during
52 an eleven month period.
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- 54 5. Nutrient leaching and uptake characteristics from three nitrogen **liquid**
55 **fertilization** formulations for **Greenhouse-Grown Azaleas** during an eleven
56 month period.
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- 58 6. Nutrient leaching and uptake characteristics from three nitrogen **liquid**
59 **fertilization** formulations for **Outdoor-Grown Ligustrum** during an eleven
60 month period.
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81 **Nutrient Release Characteristics of Four Controlled-Release Fertilizers in Acid**
82 **Substrate During an 11-Month Period in a Greenhouse Environment: I. Effects on**
83 **Leachate Electrical Conductivity, pH, and Nitrogen, Phosphorus, and Potassium**
84 **Concentrations.**

85

86 **Introduction**

87

88 The Federal Water Pollution Control Act of 1948 was overhauled in 1972 with the
89 addition of several amendments that directed regulations towards non-point
90 sources of pollution such as commercial and private landscapes and nurseries.

91 Since that time, the Act has been referred to as the Clean Water Act and, among
92 other items, indicates that all states must identify impaired waterbodies in their
93 region and must implement regulations to mitigate pollutant runoff from any point
94 and non-point sources. States are to use the U.S. Environmental Protection
95 Agency (EPA) set of guidelines referred to as the Total Maximum Daily Loads
96 (TMDLs) process described in §101(a), which provides a narrative list of physical
97 and chemical parameters that should be considered as possible pollutants.

98 Within this list are two major plant nutrients, nitrogen (N) and phosphorus (P),
99 which are used as fertilizers in the nursery industry. Other chemical parameters
100 related to the horticultural industry, such as pH and salinity, are also listed in
101 these federal guidelines.

102

103 Of the chemicals listed in these guidelines, N has the greatest risk for
104 contaminating runoff, since N usage by the container nursery industry is relatively
105 high (536 lb·A⁻¹ each year) compared to other chemicals used on horticultural
106 crops (Rathier and Frink, 1989) and various cultural practices of the industry are
107 highly conducive to nitrate (NO₃⁻) leaching. In order for the nursery industry to
108 act in accordance with these regulations, it is imperative that more efficient
109 fertilization and irrigation guidelines be developed and more effective fertilizers
110 be designed so that nutrient use efficiency (NUE) is optimized and nutrient
111 leaching is minimized.

112

113 Controlled-release fertilizers (CRFs) are used extensively for the production of
114 containerized woody ornamental plants, with the nursery and greenhouse
115 industry accounting for almost 20% of all CRFs used in the United States
116 (Goertz, 1993). Most of the research related to nutrient release characteristics
117 from CRFs have been undertaken using atypical substrates such as 100% sand,
118 under controlled laboratory conditions (Broschat, 1996; 2005; Handreck, 1997;
119 Havis and Baker, 1985a; Holcomb, 1981; Huett and Gogel, 2000; Husby et al.,
120 2003; Lamont et al., 1987; Lunt and Oertli, 1962; Oertli and Lunt, 1962; Patel and
121 Sharma, 1977), or by means of field studies (Gandeza et al., 1991), none of
122 which reflect conditions experienced in container production facilities. In
123 addition, the duration of many experiments has been limited to six months or
124 less, which is insufficient for measuring nutrient release from 12-month-release

125 fertilizer formulations, as alluded to by Havis and Baker (1985a). In the following
126 study, nutrient release patterns of four types of CRFs, when blended into an acid
127 substrate, were measured during an 11-month period in an unheated
128 greenhouse environment to determine characteristics of nutrient release under a
129 typical production scenario.

130

131 **Materials and Methods**

132

133 *Fertilizer treatments.* Treatments consisted of four different CRFs: Osmocote
134 24-4-9 (Scott-Sierra Horticultural Products Co., Marysville, Ohio), Nutricote 18-6-
135 8 Total (Chisso-Asahi Fertilizer Co., Tokyo), Multicote 17-5-11 + minors (Haifa
136 Chemicals, Ltd., Haifa Bay, Israel), and Polyon 17-5-11 + micros (Pursell
137 Technologies, Inc., Sylacauga, Ala.). All four fertilizers release nutrients based
138 on water diffusion into prills, the rate of which is regulated by temperature.
139 Osmocote did not contain micronutrients; therefore, micronutrients were provided
140 by incorporation of Micromax (Scott-Sierra Horticultural Products Co., Marysville,
141 Ohio), an uncoated, granular fertilizer at a rate of 0.53 kg·m³. All CRFs were
142 365-day release formulations. However, release rates of these CRFs are based
143 on different temperature regimes: 27°C for Osmocote and Polyon, 21°C for
144 Multicote, and 21-27°C for Nutricote. Element concentrations and compounds
145 used in each fertilizer were different (Table 1). Since the percentage of nutrients
146 contained in the different fertilizers varied, the amount of fertilizer added was
147 calculated so that all treatments contained 3.11 g N per container, which is

148 equivalent to $1.17 \text{ kg N}\cdot\text{m}^{-3}$, a rate most fertilizer companies recommended for
149 slow-growing woody ornamentals. While N content for all CRF treatments was
150 the same, the content of other nutrients varied. CRF prills were incorporated
151 throughout the substrate using a portable cement mixer. Substrate with CRF
152 was placed in #1 black polyethylene containers (2.4 L; 157 mm top diameter, 127
153 mm bottom diameter, 178 mm tall; Farrand Enterprises, Chino, Calif.) with side
154 and bottom drainage holes. Each container represented one replication. There
155 were five replications of each treatment for each of 11 months of leachate
156 monitoring. Containers were placed in a randomized complete block design.
157

158 *Leachate collection.* Leachate was collected from each container by placing a
159 plastic sleeve (140 mm bottom diameter and 165 mm top diameter) over each
160 container, each container and sleeve were placed into a plastic 2.45 L (152 mm
161 top diameter, 127 mm bottom diameter, and 152 mm high) bucket. The larger
162 dimensions of the container compared to the collection bucket allowed the
163 containers to be elevated above the level of the leachate. Leachate was
164 collected from buckets twice per week, and volumes from each bucket were
165 combined so that there was one leachate sample per container per week.
166 Leachate electrical conductivity (EC) and pH were measured once per week after
167 the first irrigation of the week. EC was measured with a Horiba conductivity
168 meter model B-173 (Horiba Ltd., Minami-ku Kyoto, Japan) and pH was measured
169 with a Horiba compact pH meter model B-213 (Horiba Ltd., Minami-ku Kyoto,
170 Japan). Immediately following EC and pH measurements, 2 ml of 2 N sulfuric

171 acid were added to each bucket to chemically stabilize the leachate. An
172 additional 2 ml of 2 N sulfuric acid was then added to the emptied bucket, so that
173 subsequent leachate collected from the container during the final part of the
174 week was immediately stabilized. Leachate from the first collection of the week
175 was stored at 4°C until the end of the week, at which time leachate from the
176 remainder of the week was added to the weekly sample. At the end of each
177 week, collection buckets were washed with 10% bleach to prevent algal growth.

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179

180 **Results and Discussion**

181

182 *Air temperature.* Weekly average greenhouse air temperature ranged between
183 24 and 30°C during the first ten weeks (Aug. through Sept. 2001) and last ten
184 weeks (May 2002 through June 2002) of the study (Figure 1). Maximum air
185 temperatures during this time period ranged from 28 to 46°C, which are
186 temperatures observed in other warm-temperate and subtropical growing regions
187 (Ingram, 1981; Lamont et al., 1987). These temperatures were consistently
188 above 21°C, the labeled substrate temperature rating for 1-year longevity for
189 Multicote. However, only weekly maximum temperatures were typically above
190 the average labeled temperature ratings of 24°C for Nutricote and 27°C for
191 Osmocote and Polyon. Since nutrient release from many CRFs, including those
192 in the present study, is temperature dependent, any elevated temperatures that
193 may occur can greatly impact nutrient release characteristics of CRFs. In studies

194 related to environmental conditions and CRFs, nutrient release rates increased
195 by up to 200% for every 10°C increase above optimum release temperature
196 (Husby et al., 2003; Kochba et al., 1990; Lamont et al., 1987; Oertli and Lunt,
197 1962); however, other experiments (Huett and Gogel, 2000) showed only a 15%
198 increase in release rates at higher temperatures. At suboptimum temperatures,
199 nutrient release characteristics have been shown to be inconsistent (Engelsjord
200 et al., 1997); however, others (Kochba et al., 1990) have shown that the nutrient
201 release rates are lower, but uniform.

202

203 *Electrical conductivity.* Leachate electrical conductivity (EC) was elevated during
204 the first five months of the study, relative to the later time frame, regardless of
205 fertilizer type, with significant differences observed among treatments (Figure 2).
206 Similar trends in EC fluctuations have also been observed in other studies
207 (Lamont et al., 1987). During the entire study, the leachate EC of the Osmocote
208 treatment was often significantly lower than the EC from the other treatments,
209 especially Multicote. Both Polyon and Nutricote treatments also had leachate EC
210 levels that were significantly lower than levels recorded for Multicote. Near the
211 end of the study, (weeks 45 and 46), the leachate EC for Osmocote was
212 significantly greater than the EC for Multicote. In diurnal studies (20 hours) that
213 compared Osmocote, Nutricote, and Polyon, Osmocote resulted in leachates
214 with lower soluble salts than Nutricote or Polyon at temperatures between 20 and
215 25°C; however, Osmocote had higher release rates at temperatures above 35°C
216 (Husby et al., 2003). In longer-term studies (Lamont et al., 1987), Osmocote

217 exhibited higher release rates than Nutricote at all temperature regimes studied
218 (5, 10, 15, 20, 25, 30, 35, 40, and 45°C). Based on the present study and
219 research by Husby et al. (2003) and Lamont et al. (1987), there are daily as well
220 as weekly trends in nutrient release from CRFs which, in the case of diurnal
221 fluctuations, closely reflect changes in temperature. However, over extended
222 periods of time, the pool of nutrients remaining in the prills diminishes, which
223 results in decreased quantities of nutrient release and thus decreased EC levels,
224 even with increased temperatures during later parts of a production period.

225

226 From an environmental perspective, federal guidelines dictate that soluble salts
227 should not exceed 250 mg·L⁻¹, based on chloride and sulfate concentrations
228 only. Therefore, monitoring and control of EC, from a nursery production point-
229 of-view, would be irrelevant, since EC takes into account all dissolved salts.

230

231 *pH*. During the first 20 weeks, leachate pH was variable, but consistently acidic
232 (Figure 3), even though irrigation water pH was 7.5, indicating that substrate
233 and/or fertilizers used in the study influenced leachate pH more so than did the
234 neutral irrigation water. Other studies (Chen et al., 2003) have demonstrated
235 that a similar substrate containing pine bark, sand, and peat buffered irrigation
236 water of pH 9.7 (with low alkalinity) to a measured substrate pH of approximately
237 6.5. In addition, a study by Ivy et al. (2002) showed that fertilizer type (Osmocote
238 and Polyon) used does not influence pH.

239

240 Among fertilizer types, few significant differences were detected. Leachate pH
241 from substrate containing Osmocote was higher compared to the leachate pH of
242 Multicote and Nutricote treatments (two and four times, respectively) during the
243 47-week period. The pH readings in weeks 42 and 43 for Multicote and week 42
244 for Nutricote were significantly higher than the leachate pH for Osmocote.
245 However, another study (Ivy et al., 2002) showed that fertilizer type (Osmocote
246 and Polyon) did not influence pH but, in a different study (Argo and Biernbaum,
247 1997), it was determined that leachate pH is influenced by compounds used in
248 fertilizers, especially in respect to the ratios of NH_4^+ to NO_3^- , with pH decreasing
249 when the amount of NH_4^+ relative to NO_3^- increases. In the present study, the
250 amount of N as NH_4^+ ranged between 43% and 53%, except for the Osmocote
251 formulation that contained 76% of the N in the ammoniacal form (NH_4^+ and urea).
252 These differences in N form may not have been different enough to cause
253 significant changes in leachate pH among different fertilizer treatments. None of
254 the changes in pH measured in the present study fell outside the critical ranges
255 established by the U.S. EPA (pH of 5.0-9.0 for domestic water supplies, 6.5 - 8.5
256 for freshwater aquatic life, and 6.5 – 9.0 for marine aquatic life) (U.S. EPA, 1976).

257

258 *Ammoniacal-nitrogen concentration.* During the first four weeks of the study,
259 NH_4^+ -N concentrations in leachates in all treatments were below $10 \text{ mg}\cdot\text{L}^{-1}$
260 (Figure 4). From week 5 to week 9, concentrations increased to above 150
261 $\text{mg}\cdot\text{L}^{-1}$, followed by a decline from week 10 to week 30. During the last 17
262 weeks, leachate NH_4^+ -N concentrations were less than $1 \text{ mg}\cdot\text{L}^{-1}$, except for the

263 Osmocote treatment, which averaged ~3.0 mg during the last three weeks of the
264 study, which was significantly greater than with the other fertilizer treatments.
265 When making comparisons among treatments, leachate $\text{NH}_4^+\text{-N}$ from the
266 Multicote treatment was significantly greater than with the other treatments
267 several times during the first 17 weeks of the study. Similar release patterns
268 were observed in other experiments (Broschat, 2005) using 8- to 9-month-
269 release formulations of Osmocote and Nutricote. Broschat (2005) found that
270 30% to 50% of the $\text{NH}_4^+\text{-N}$ was released from prills of these fertilizers by the end
271 of the second month, and by the seventh month, less than 20% of the $\text{NH}_4^+\text{-N}$
272 remained in the prills. While patterns of $\text{NH}_4^+\text{-N}$ leaching were similar to release
273 rates measured by Broschat (2005), the percentage of $\text{NH}_4^+\text{-N}$ from the CRF
274 recovered in the leachates in the present study were lower, which is probably
275 associated with the binding of $\text{NH}_4^+\text{-N}$ to organic matter of the container
276 substrate, as described by others (Foster et al., 1983; Thomas and Perry, 1980).
277 The decrease in $\text{NH}_4^+\text{-N}$ concentrations of leachates during the later stages of
278 the present study may be partly accounted for by lower quantities of $\text{NH}_4^+\text{-N}$ that
279 remained in the prills, as was determined by Broschat (2005). In addition, there
280 was probably nitrification of NH_4^+ to NO_3^- in the present study. In studies
281 associated with nitrification of NH_4^+ , it was determined that the majority of the
282 NH_4^+ is converted to NO_3^- , with a slower rate of nitrification when soil pH was
283 lower (Dancer et al., 1973; Niemiera and Wright, 1986) or when substrate
284 temperatures exceed 46°C (Walden and Wright, 1995). In a diurnal study
285 (Husby et al., 2003), $\text{NH}_4^+\text{-N}$ release was closely associated with temperature,

286 with greater release of NH_4^+ from Osmocote compared to Nutricote or Polyon at
287 high temperatures ($>35^\circ\text{C}$), but lower release rates at low temperatures ($<35^\circ\text{C}$).
288 In the present study, air temperatures were seldom above 35°C , which may
289 explain the fairly level $\text{NH}_4^+\text{-N}$ release rates.

290

291 Federal guidelines do not mention $\text{NH}_4^+\text{-N}$ concentration, but do state that
292 concentrations of un-ionized ammonia (NH_3), expressed as NH_3 , should not
293 exceed $0.02 \text{ mg}\cdot\text{L}^{-1}$ for freshwater aquatic life (U.S. EPA, 1976). Based on other
294 research (Thurston et al., 1974), this would mean that total ammonia ($\text{NH}_4^+ +$
295 NH_3) concentration, which was measured in the present study, would have to be
296 in the range of 16 to $160 \text{ mg}\cdot\text{L}^{-1}$ for a solution pH range of 7.0 to 6.0, but is also
297 dependent on solution temperature. Given these criteria, it is likely that NH_3
298 concentrations could have exceeded federal limits several times throughout the
299 course of the experiment.

300

301 *Nitrate-nitrogen concentration.* In all fertilizer treatments, $\text{NO}_3^-\text{-N}$ concentrations
302 were below $10 \text{ mg}\cdot\text{L}^{-1}$ during the first four weeks of the study, with the exception
303 of week 2, when leachate from the Osmocote treatment averaged $32 \text{ mg}\cdot\text{L}^{-1}$
304 (Figure 5). From week 4 to week 9, $\text{NO}_3^-\text{-N}$ concentrations increased to over 100
305 $\text{mg}\cdot\text{L}^{-1}$ for all treatments except Osmocote. From week 10 to the end of the
306 study, $\text{NO}_3^-\text{-N}$ concentration decreased and leveled off to approximately 30
307 $\text{mg}\cdot\text{L}^{-1}$ for leachates from treatments containing Nutricote, Polyon, or Multicote.
308 However, for Osmocote, leachate $\text{NO}_3^-\text{-N}$ concentrations gradually increased to

309 approximately 50 mg·L⁻¹ and then decreased to about 30 mg·L⁻¹ during the
310 remaining 20-week period. Of the fertilizers tested, Osmocote appeared to
311 produce the most stable release rate of NO₃⁻-N. During one-third of the weeks of
312 the study, NO₃⁻-N concentrations were lower for Osmocote compared with the
313 leachate collected from the other fertilizer types. Similarly, the NO₃⁻-N
314 concentration in leachates collected from the Polyon treatment was significantly
315 lower than for Multicote six weeks out of ten during weeks 25 through 35. The
316 release pattern of NO₃⁻ in leachates from Osmocote and Nutricote treatments are
317 similar to the release characteristics determined by others (Broschat, 2005;
318 Prasad and Woods, 1971), except that, on a weekly basis, the data from the
319 present study indicated that Osmocote may have a lower, but more stable, NO₃⁻
320 release pattern from the prills compared to Nutricote. In general, the elevated
321 concentrations of NO₃⁻ relative to NH₄⁺, especially during the later half of the
322 study, are probably associated with the nitrification of NH₄⁺ in addition to the
323 release of NO₃⁻ from the fertilizer prills. Other studies have demonstrated
324 significant nitrification of NH₄⁺ in soils and substrates of acidic to neutral pH
325 (Dancer et al., 1973; Niemiera and Wright, 1986).

326

327 Based on federal guidelines of a maximum of 10 mg·L⁻¹ of NO₃⁻-N for domestic
328 water use (U.S.EPA, 1976), NO₃⁻ -N concentrations exceeded permissible levels
329 during most of the experiment. However, if plants had been included in the
330 present study, the likelihood of NO₃⁻ leaching in excess of 10 mg·L⁻¹ would
331 probably only exist during the first 12 weeks of the study when total

332 concentrations of NO_3^- -N were above $100 \text{ mg}\cdot\text{L}^{-1}$. The only fertilizer product in
333 exception to this was Osmocote, which had relatively low NO_3^- -N concentrations
334 during the entire period of the study.

335

336 **Total inorganic-N concentrations.** Release characteristics as indicated by total
337 inorganic-N concentrations (ammonium + nitrate) showed significant release of N
338 during the first nine weeks of the study, then a decrease and leveling off period
339 during the remaining 36 weeks for Polyon, Nutricote, and Multicote. The
340 exception to this pattern was Osmocote, which had more consistent leachate
341 inorganic N concentrations between 30 and $50 \text{ mg}\cdot\text{L}^{-1}$ throughout the entire
342 experimental period. The pattern of high then low N release rates during the
343 beginning and end of the experiment, respectively, are similar to other studies
344 (Huett, 1997b; Huett and Gogel, 2000; Prasad and Woods, 1971; Patel and
345 Sharma, 1977). The initial lag in N appearance in leachate in the present study
346 may be also accounted for by the chemical and biological immobilization of N in
347 the organic substrate, which has been documented by others (Foster et al., 1983;
348 Gartner et al., 1971; Handreck and Bunker, 1996; Thomas and Perry, 1980). In
349 addition, concentrations of NH_4^+ relative to NO_3^- were similar during the first half
350 of the study, but NO_3^- was the predominant inorganic N form during the later half
351 of the study, regardless of fertilizer type. This phenomenon is probably
352 associated with the nitrification of NH_4^+ , as demonstrated in other studies
353 (Dancer et al., 1973) and possibly the faster release rate of NH_4^+ -N from the
354 prills, relative to NO_3^- , as measured by Broschat (2005). In shorter-term studies

355 (6 months), NO_3^- was the predominant N form during the entire experimental
356 period (Cabrera, 1997). Regardless of fertilizer type, the present study did not
357 detect any significant correlation of inorganic N release rate with air
358 temperatures. These results differ from studies conducted by Cabrera (1997)
359 and Handreck (1997), where N release characteristics of Osmocote, Nutricote,
360 and Polyon were closely associated with temperature. However, the
361 experiments conducted by Cabrera lasted 9 months, during which time
362 temperatures slowly increased then decreased, and the studies conducted by
363 Handreck were in a temperature-controlled (21°C) laboratory. In the present
364 study, temperatures fluctuated, with relatively high temperatures during the first 2
365 months, then decreasing during the winter, and then increasing again during the
366 following spring.

367

368 *Total phosphorus concentrations.* Since fertilizer additions were normalized for N
369 content, total P content added differed among fertilizer treatments, with an
370 average P content of 0.40 g/container for Polyon, Nutricote, and Multicote, but
371 50% less for Osmocote (0.22 g/container). P concentrations in leachates of all
372 treatments fluctuated throughout the study, but were higher (15 to > 60 $\text{mg}\cdot\text{L}^{-1}$)
373 during the first 10 weeks compared to the last 27 weeks of the study, when
374 concentrations averaged below 10 $\text{mg}\cdot\text{L}^{-1}$ (Figure 7). The pattern of high then
375 low P release rates in the present study is similar to other studies with Osmocote,
376 Nutricote, and Polyon (Huett, 1997b; Huett and Gogel, 2000). When comparing
377 treatments, Multicote tended to produce the highest P concentrations in

378 leachates compared to the other fertilizers types, primarily during the first 37
379 weeks. For the most part, total P concentrations of other treatments were
380 somewhat similar to each other throughout the study. In another experiment
381 (Broschat, 2005), release of P from Osmocote prills during the first five months
382 was significantly greater than that measured from Nutricote; however, by the
383 seventh month, differences in release were not significant. In shorter-term
384 studies (10 weeks), there were no differences in P release rates between
385 Osmocote and Nutricote (Huett, 1997b). According to field and container
386 experiments (Flint, 1962; Handreck, 1996; Havis and Baker, 1985a; 1985b;
387 Wright, 1984; Yeager and Wright, 1982), solution P concentrations of
388 approximately $10 \text{ mg}\cdot\text{L}^{-1}$ are required for optimal growth of woody ornamentals.
389 A fraction of this P is tied in organic substrates, as demonstrated by Handreck
390 (1996). Based on the present study and the experiments of others, it appears
391 that a high release of P from CRFs may elevate the risk of P leaching during the
392 early parts of the production cycle, but a lower release of P may limit plant growth
393 during the later part of a production cycle.

394

395 Regarding P, the U.S. EPA has only established a concentration limit of 0.01
396 $\mu\text{g}\cdot\text{L}^{-1}$ for elemental P (U.S. EPA, 1976). There are currently no federal
397 guidelines for ortho-phosphate or total phosphorus. Based on the criterion for
398 total P, the greatest risk of total P leaching would be during the first 20 weeks of
399 the production cycle, especially if any fraction of the total P determined was
400 elemental in nature.

401

402 *Total potassium concentrations.* Total K concentrations of leachates from all
403 treatments ranged between 10 and 170 mg·L⁻¹, with higher concentrations during
404 the first 20 weeks of the study compare to the last 27 weeks (Figure 8). When
405 comparing CRF types, K concentrations in leachates from Osmocote and
406 Nutricote treatments were similar, but both were significantly lower than K
407 concentrations of leachates collected from containers containing Polyon or
408 Multicote. These differences may be attributed to the amount of K added, since
409 fertilizer additions were normalized for N. Osmocote and Nutricote contained
410 approximately 36% less K than the Polyon and Multicote treatments. While not
411 measured, a portion of the K released from the prills was probably tied up by the
412 organic fraction, which has been shown to occur with organic substrate (Brown
413 and Pokorny, 1977; Foster et al., 1983). Elevated K release from CRFs during
414 the early part of the production cycle observed in the present study is similar to
415 other studies (Broschat, 1996, 2005; Huett, 1997b) that compared K release from
416 Osmocote, Nutricote, and Multicote; however, Broschat (1996, 2005) did not
417 detect differences among fertilizer types. In other research (Holcomb, 1981) with
418 Osmocote, K release rates were linear throughout the 63-day study period. Even
419 though K concentrations were elevated in leachates during the first portion of the
420 experiment, there is no concern of K concentrations exceeding federal
421 recommended guidelines, since no guidelines have been established for K as of
422 yet.

423

424 **Conclusions**

425

426 Under greenhouse conditions, where high temperatures were moderated with
427 evaporative cooling pads and fans, the release characteristics of NH_4^+ , NO_3^- , K,
428 and P from all CRFs tested were elevated during the first half of the study,
429 followed by lower release rates during the later half of the 11-month period, even
430 when temperatures increased during the spring. Differences were noted among
431 fertilizer types. Patterns of leachate EC reflect somewhat the leachate
432 concentrations of N, P, and K. These correlations have also been measured by
433 others (Cabrera, 1997; Huett and Morris, 1999; Husby et al., 2003). Leachate
434 EC is probably associated with both nutrient release from CRF and soluble salts
435 leached out of the substrate during first few weeks of the study. Based on the
436 results of this study and data from other long term studies, it appears that nutrient
437 release from CRFs may be in excess of plant requirements during the first half of
438 the production period (Huett, 1997b; Huett and Gogel, 2000), but may be
439 insufficient during later stages of production (Huett, 1997a), depending on the
440 nutrient demands of the crop being grown and the temperature profiles during
441 production.

442

443 From an environmental perspective, risk of water impairment, when using the
444 CRFs currently studied, would be greatest during the first 20 weeks of crop
445 production, since EC, $\text{NH}_4\text{-N}$, $\text{NO}_3\text{-N}$, total P, and total K were elevated during
446 this time period. Based on guidelines established by the U.S. EPA (1976), the

447 chemical parameters of pH, EC, $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$, and P, as observed in the
448 present study, were not within acceptable limits during various time frames
449 throughout the 11-month period. However, concentrations of nutrients in
450 leachates would probably be reduced if plants were present in the containers.
451 For this reason, additional research regarding leaching characteristics of CRFs is
452 required, particularly for a typical plant production scenario using woody
453 ornamentals commonly grown in the nursery industry.

454

455

456

457

458

459 **Nutrient Release Characteristics of Four Controlled-Release Fertilizers in Acid**
460 **Substrate During an 11-Month Period in a Greenhouse Environment II. Effects on**
461 **Leachate Concentrations of Calcium, Magnesium, Iron, Manganese, Zinc, Copper,**
462 **and Molybdenum**

463

464 **Introduction**

465

466 In addition to the two well-known pollutants, nitrate (NO_3^-) and phosphate (PO_4^{2-}),
467 other essential plant nutrients listed in §101(a) of the Clean Water Act (U.S.
468 Environmental Protection Agency, 1994) are boron (B), copper (Cu), iron (Fe),
469 manganese (Mn) and zinc (Zn). Although not receiving as much attention as
470 NO_3^- and PO_4^{2-} , these plant micronutrients are chemicals that are used by the

471 nursery industry and, therefore, are at risk of being regulated by state agencies.
472 The likelihood of micronutrient contamination of surface waters from nursery
473 runoff is unknown. Therefore, information regarding the movement of these
474 nutrients in nursery production systems is needed so that the nursery industry is
475 aware of any potential problems associated with micronutrient fertilization so that
476 Best Management Practices (BMPs) may be developed to mitigate the runoff of
477 these nutrients.

478

479 Controlled-release fertilizers (CRFs) are commonly used in the nursery industry.
480 When used properly, these types of fertilizers can reduce the amount of nutrients
481 in runoff compared to water-soluble fertilizers. However, little information is
482 available regarding leaching of plant essential nutrients other than N and P. In
483 the following study, the nutrient release patterns of four types of CRFs, when
484 blended into a substrate, were measured during an 11-month period in an
485 unheated greenhouse environment to determine characteristics of nutrient
486 release under a typical production scenario.

487

488 **Materials and Methods**

489

490 The methodology of the experiment has been described in the first manuscript of
491 this series. Therefore, methods are briefly described here, with emphasis on
492 application, measurement and analyses of calcium (Ca), magnesium (Mg), Fe,
493 Mn, Zn, Cu, and Mo.

494

495 *Substrate.* Substrate consisted of 5 sphagnum peatmoss (Premier Horticulture
496 Inc., Red Hill, Penn.) : 4 pine bark (6.4-9.5 mm) : 1 washed builders sand (by
497 volume). Substrate was amended with dolomite 65 (Chemical Lime Co.,
498 Scottsdale, Ariz.) at a rate of $0.59 \text{ kg}\cdot\text{m}^{-3}$ and ultrafine calcium sulfate (Western
499 Mining and Minerals, Apex, Nev.) at a rate of $0.59 \text{ kg}\cdot\text{m}^{-3}$. Substrate and
500 amendments were blended together using a Model MB20L5 Batch Mixer
501 (Measured Marketing, Kankakee, Ill.). Substrate was analyzed for nutrient
502 concentrations prior to incorporation of CRFs by grinding substrate to pass
503 through a 40-mesh screen. The ground sample was then extracted by adding
504 100 ml deionized water to 50 ml volume of ground substrate and then filtered.
505 Nutrient concentrations in extracts were quantitatively determined using an
506 inductively coupled plasma (ICP) spectrometer (Thermo Electron Corp., model
507 IRIS 1000 HR, Franklin, Mass.) Substrate nutrient concentrations prior to
508 incorporation of CRFs were as follows ($\text{mg}\cdot\text{L}^{-1}$): 13.06 Ca, 10.19 Mg, 2.62 Fe,
509 0.59 Mn, 0.75 Zn, 0.11 Cu, 0.01 Mo.

510

511 *Fertilizer treatments.* Treatments consisted of four different types of CRFs:
512 Osmocote 24-4-9 (Scott-Sierra Horticultural Products Co., Marysville, Ohio),
513 Nutricote 18-6-8 Total (Chisso-Asahi Fertilizer Co., Tokyo), Multicote 17-5-11 +
514 minors (Haifa Chemicals, Ltd., Haifa Bay, Israel), and Polyon 17-5-11 + micros
515 (Pursell Technologies, Inc., Sylacauga, Ala.). All four fertilizers release nutrients
516 based on water diffusion into prills, the rate of which is regulated by temperature.

517 Osmocote did not contain micronutrients; therefore, micronutrients were provided
518 by the incorporation of Micromax (Scott-Sierra Horticultural Products Co.,
519 Marysville, Ohio). All CRFs were 365-day release formulations. However,
520 release rates of the different products are based on different temperature
521 regimes: 27°C for Osmocote and Polyon, 21°C for Multicote, and 21-27°C for
522 Nutricote. Element concentrations and the compounds used in each fertilizer
523 were different (Table 1). Since the percentage of nutrients contained in the
524 different fertilizers varied, the amount of fertilizer added was calculated so that all
525 treatments contained 3.11 g N per container, which is equal to 1.17 kg N·m⁻³, a
526 rate most fertilizer companies recommend for slow-growing woody ornamentals.
527 While N content for all CRF treatments were the same, the content of other
528 nutrients varied. The CRF prills were incorporated throughout the substrate
529 using a portable cement mixer. Substrate with CRF were placed in #1 black
530 polyethylene containers (2.4 L; 157 mm top diameter, 127 mm bottom diameter,
531 178 mm tall; Farrand Enterprises, Chino, Calif.) with side and bottom drainage
532 holes. Each container represented one replication. There were five replications
533 of each treatment for each month of leachate monitoring. Containers were
534 placed in a randomized complete block design.

535

536 *Cultural Practices.* Substrate and fertilizer treatments were prepared on 1 Aug.
537 2001. The experiment was conducted from 1 Aug. 2001 through 27 June 2002.
538 Containers were placed in an unheated, poly-covered greenhouse on the
539 University of California, Riverside campus (lat. 33°53'30"N, long. 117°15'00"W).

540 Containers were irrigated with potable water, which contained ($\text{mg}\cdot\text{L}^{-1}$): 64.30 Ca,
541 9.48 Mg, 0.01 Fe, 0.00 Mn, 0.00 Cu, 0.004 Zn, and 0.003 Mo. Irrigation water
542 was distributed uniformly over the surface of the substrate using DRT4-36 ring
543 drippers (Dramm Corporation, Manitowoc, Wis.). Containers were irrigated
544 approximately every 2 d, providing an average irrigation volume of ~ 1 L/week
545 and an average leachate volume of ~ 750 ml/ week.

546

547 *Leachate collection.* All leachate was captured from each container, collected
548 twice per week, and volumes from each collection bucket were combined so that
549 there was one leachate sample per container per week. 2 ml of 2 N sulfuric acid
550 were added to each collection bucket to chemically stabilize the leachate.

551

552 **Results**

553

554 *Calcium concentration.* During the first 13 weeks of the experiment, Ca
555 concentrations in leachates from all treatments fluctuated, with concentrations
556 reaching as high as $300 \text{ mg}\cdot\text{L}^{-1}$ (Figure 1). During the last 30 weeks of the study,
557 Ca concentrations in leachates ranged between 25 and $50 \text{ mg}\cdot\text{L}^{-1}$, with few
558 differences among treatments. During certain weeks, some treatments were
559 significantly different from others; however, there did not appear to be any
560 discernable pattern or trend in these differences. The lack of differences among
561 treatments is expected since most of the Ca in all treatments was derived from
562 calcium sulfate and dolomite that were blended into the substrate in addition to

563 Ca present in the potable water. Ca concentrations observed in the present
564 study are similar to leachate readings observed in simulated production systems
565 (Chen et al., 2003), where Ca concentrations ranged from 12 to 42 mg·L⁻¹. In
566 shorter-term (1-2.5 months) studies (Huett, 1997b; Huett and Morris, 1999), the
567 greatest loss of Ca occurred during the first week when testing Osmocote and
568 Nutricote.

569

570 Based on other research, it does not appear that Ca would have been limiting to
571 plant growth during any period of the study. In hydroponically-grown New
572 Guinea impatiens (*Impatiens* 'Equinox'), maximum Ca uptake rate was achieved
573 at a Ca concentration of approximately 95 mg·L⁻¹ (Mankin and Fynn, 1996).
574 However, in several other studies (Dunham and Tatnall, 1961; Edwards and
575 Horton, 1981; Starr and, 1984; Wright, 1984), optimum plant growth was
576 obtained with a Ca leachate concentration of 10-15 mg·L⁻¹. If similar uptake
577 kinetics and Ca requirements can be generalized for most woody ornamentals,
578 then sufficient Ca, under the fertilization regime described, should be available
579 throughout the production period.

580

581 From an environmental perspective, there are no federal guidelines established
582 for Ca. Significant Ca runoff may indirectly contribute to elevated electrical
583 conductivity and pH. However, EPA criterion for salinity is 250 mg·L⁻¹ and is
584 based only on chloride and sulfates (U.S. EPA, 1976), not Ca or any other
585 element. Present EPA criteria for pH are 5.0 to 9.0 (domestic water supply), 6.5

586 to 9.0 (freshwater aquatic life), and 6.5 to 8.5 (marine aquatic life) (U.S. EPA,
587 1976). The other water quality criterion associated with Ca is alkalinity, which is
588 not to exceed 20 mg·L⁻¹ of calcium carbonate (CaCO₃) for freshwater aquatic life,
589 “except where natural concentrations are less”. (U.S. EPA, 1976).

590

591 *Magnesium concentrations.* Leachate concentrations of Mg fluctuated during the
592 first 12 weeks of the study, with concentrations as low as 4 mg·L⁻¹ for Osmocote
593 (week 7) and as high as 70 mg·L⁻¹ for Nutricote (week 9) (Figure 2). However,
594 from week 13 to the conclusion of the experiment, Mg concentrations were
595 relatively stable, falling within the range of 10-20 mg·L⁻¹. The range of Mg
596 concentrations observed in the present study are similar to concentrations
597 measured in simulated container production systems (Chen et al., 2003). In
598 shorter-term studies (Huett, 1997b; Huett and Morris, 1999), Mg leaching was
599 also greatest during the first week of the experiment. Several significant
600 differences were measured among treatments, but there was no noticeable
601 pattern in these differences. Other studies noted that Mg in the form of sulfates
602 solubilized and leached from substrate more quickly than Mg derived from oxides
603 and carbonates (Broschat and Donselman, 1985).

604

605 Based on plant requirements for Mg that have been established in other studies
606 (Mankin and Fynn, 1996; Starr and Wright, 1984; Wright, 1984), it appears that
607 Mg availability would be sufficient for plant growth in the production scenario
608 studied. In hydroponically grown herbaceous plants, the maximum Mg uptake

609 rate was achieved at a concentration of 30 mg·L⁻¹ (Mankin and Fynn, 1996). In
610 other studies (Starr and Wright, 1984; Wright, 1984), optimum plant growth in
611 containerized plants was achieved when leachate Mg concentrations were
612 maintained between 10 and 15 mg·L⁻¹. If the research by Mankin and Fynn
613 (1996) and Wright (1984) can be applied to general containerized production
614 practices, Mg availability from CRFs should not be limiting during a typical 11-
615 month plant production cycle.

616

617 Environmentally, the greatest likelihood of Mg leaching from containers would
618 only be during the first ten weeks of a production cycle; thereafter, plant roots
619 should take up all Mg. However, since current federal policies (U.S. EPA, 1976)
620 have no criterion for Mg concentration in runoff waters, Mg in runoff is not
621 presently an environmental concern.

622

623 *Iron concentrations.* Iron concentrations in leachates from all treatments were
624 relatively high and fluctuated during the first 12 weeks of the study (Figure 3). Fe
625 concentrations in leachates were relatively stable during the last 30 weeks of the
626 study, with significantly higher concentrations recovered in the leachates from the
627 Nutricote treatment compared to the other three CRF treatments. This may be
628 due to Fe leaching out of broken prills or, in the case of Micromax, Fe directly
629 solubilizing from the fertilizer since Micromax is a non-coated form of
630 micronutrient fertilizer. In other studies with uncoated granular fertilizers, a
631 notable leaching of Fe occurred (Frost et al., 2003; Handreck, 1989) and, in

632 some cases, the leaching rate from granular fertilizers was greater than that
633 observed from soluble fertilizers (Frost et al., 2003). Overall, Fe concentrations
634 measured in the present study, with the exception of the first 11 weeks, were
635 similar to those observed in containerized production systems (Chen et al.,
636 2003). The significant differences with Nutricote may be associated with the iron
637 formulation, which was iron EDTA, a soluble iron form. Iron chelate was also the
638 form of Fe used with Multicote; however, the Multicote treatment resulted in
639 significantly greater leaching than all other CRFs during the first three weeks of
640 the study, which may have resulted in less Fe available for release during the
641 later part of the study. Broschat and Donselman (1985) noted that the leaching
642 of Fe from organic-based substrate was greater with the use of chelated Fe
643 compared to other mineral forms of Fe. It has also been determined that Fe is
644 immobilized by organic substrate, up to $100 \text{ mg}\cdot\text{L}^{-1}$ for certain types of sawdust
645 and bark products (Handreck, 1989).

646

647 Based on the present study and the results of other research (Broschat and
648 Donselman, 1985; Chen et al., 2003; Frost et al., 2003, Handreck, 1989), it
649 appears that sufficient Fe was probably available for normal plant growth in the
650 simulated production system studied, with the environmental risk of high Fe
651 concentrations ($>0.3 \text{ mg}\cdot\text{L}^{-1}$) in leachates occurring during the first 20 weeks of
652 the production cycle. The primary concern for Fe leaching would be with
653 formulations using chelated iron, which have been found to be easily leached
654 from organic substrates typically used in many containerized production systems.

655 Current federal guidelines indicate Fe concentrations are not to exceed 0.3
656 mg·L⁻¹ for water for domestic use and 1.0 mg·L⁻¹ for freshwater aquatic life.
657

658 *Manganese concentrations.* Manganese concentrations in leachates of all
659 fertilizer treatments were relatively high (between 1.0 and 9.0 mg·L⁻¹), but
660 variable during the first twelve weeks of the study (Figure 4). During the last 30
661 weeks of the experiment, Mn concentrations were below 2.0 mg·L⁻¹ for all
662 treatments. In other studies, Mn leached from granular fertilizers, even more so
663 than with the use of liquid fertilizers (Frost et al., 2003). The only significant
664 differences observed were that the Osmocote treatment resulted in higher Mn
665 concentrations in leachates compared to the other CRFs during the first two
666 weeks of the study, then again from weeks 36 through 39 for Polyon and
667 Multicote. Other differences were observed among fertilizer types throughout the
668 experimental period, but no patterns or trends in these differences evolved. The
669 lack of differences is not unexpected, since all fertilizer types contained
670 manganese sulfate as their form of Mn. In other studies (Broschat and
671 Donselman, 1985), it was shown that Mn might precipitate with other compounds
672 at a substrate pH above 6.2; however, in the present study, the initial substrate
673 pH was 4.5. Therefore, solubilization and possible leaching of Mn probably
674 occurred during the beginning of the study. However, as leachate pH increased,
675 and probably substrate pH as well, the solubilization, and therefore the leaching,
676 of remaining Mn was minimized. Based on other research (Tinus and McDonald,
677 1979), a Mn concentration of 0.5 mg·L⁻¹ is sufficient for healthy plant growth.

678 Therefore, the Mn concentrations observed in the present study should be
679 adequate for plant requirements. Environmentally, Mn leaching, and therefore
680 the potential for excess Mn in runoff, was high throughout the entire study since
681 federal guidelines indicate that Mn concentrations are not to exceed $0.050 \text{ mg}\cdot\text{L}^{-1}$
682 for domestic waters and $0.100 \text{ mg}\cdot\text{L}^{-1}$ "for protection of consumers of marine
683 mollusks" (U.S. EPA, 1976), levels which are below concentrations measured in
684 the present study.

685

686 *Zinc concentrations.* Leachate concentrations of Zn were relatively high and
687 variable during the first twelve weeks of the study (Figure 5). After week 13, Zn
688 concentrations were below $0.5 \text{ mg}\cdot\text{L}^{-1}$, regardless of fertilizer type. These
689 readings were higher than leachate measurements conducted in other studies
690 (Chen et al, 2003) where Zn concentrations ranged between 0.01 and 0.04
691 $\text{mg}\cdot\text{L}^{-1}$. Leaching of Zn with granular micronutrients was also noted in other
692 studies (Frost et al., 2003). Leachate Zn concentrations from the Nutricote
693 treatment were significantly higher than with other fertilizer treatments during
694 most of the study, despite the fact that the amount of Zn was three to four times
695 less in the Nutricote treatment compared to the other CRF treatments. For plant
696 growth, the Zn concentrations observed would probably be sufficient for most
697 crops, as noted in other studies (Carroll and Loneragan, 1969; Tinus and
698 McDonald, 1979). From an environmental perspective, Zn concentrations were
699 usually below the critical limit of $5.000 \text{ mg}\cdot\text{L}^{-1}$ for domestic water supplies as
700 established by the EPA.

701

702 *Copper concentrations.* Copper concentrations in leachates were usually below
703 1.00 mg·L⁻¹, regardless of fertilizer type (Figure 6). During weeks 14 and 15,
704 concentrations of Cu increased for all treatments, then decreased and stabilized.
705 There were some differences among treatments, most notable being the higher
706 Cu concentrations with Nutricote relative to the other CRFs for eight weeks out of
707 the 47-week period. In another study (Broschat and Donselman, 1985), copper
708 concentrations in leachates were relatively stable throughout an 18-month
709 period. The low Cu concentrations observed in the present research and other
710 studies may be attributed to the high affinity of Cu for organic matter (Schnitzer
711 and Skinner, 1966). The EPA has established a Cu limit of 1.0 mg·L⁻¹ for
712 domestic water supplies (U.S. EPA, 1976). Based on the current study, the Cu
713 levels resulting from fertilization should not be of concern.

714

715 *Molybdenum concentrations.* Molybdenum concentrations were variable during
716 the first 15 weeks of the study, with significantly greater concentrations measured
717 in the Nutricote treatment compared to the other fertilizer types. After week 15,
718 Mo concentrations were near 0.0 mg·L⁻¹ for all treatments. There are currently
719 no federal guidelines established for Mo in surface waters.

720

721 **Conclusions.** Concentrations of Ca, Mg, Fe, Mn, Zn, Cu and Mo in leachates
722 were relatively high during the first 10 to 16 weeks of the 11-month production
723 cycle under cultural conditions typically used for low-nutrient requiring crops such

724 as azalea and camellia. In most cases, leachate concentrations of all nutrients
725 appeared to be at levels that would be considered sufficient for healthy plant
726 growth. From an environmental perspective, only Fe and Mn were at
727 concentrations that exceeded U.S. EPA guidelines, and these elevated
728 concentrations only occurred during the first two months of the experiment.
729 Based on these results, the use of Ca, Mg, Zn, Cu and Mo in the fertilizer
730 program tested should have little or no impact on water quality of nursery runoff.
731 However, the use of Fe and Mn should be carefully considered, as
732 concentrations of these nutrients may exceed federal guidelines, especially
733 during the early phase of a typical production cycle.

734

735 **Literature Cited**

736

737 Argo, W.R. and J.A. Biernbaum. 1997. The effect of root media on root-zone pH,
738 calcium, and magnesium management in containers with impatiens. J. Amer.
739 Soc. Hort. Sci. 122:275-284.

740

741 Broschat, T.K. 1996. Release rates of soluble and controlled-release potassium
742 fertilizers. HortTechnology 6:128-131.

743

744 Broschat, T.K. 2005. Rates of ammonium-nitrogen, nitrate-nitrogen, phosphorus,
745 and potassium from two controlled-release fertilizers under different substrate
746 environments. HortTechnology 15:332-335.

747

748 Broschat, T.K. and H.M. Donselman. 1985. Extractable Mg, Fe, Mn, Zn, and Cu
749 from a peat-based container medium amended with various fertilizers. J. Amer.
750 Soc. Hort. Sci. 110:196-200.

751

752 Brown, E.F. and F.A. Pokorny. 1977. Potassium distribution and retention in pine
753 bark and sand media. HortScience 12:343-344.

754

755 Cabrera R.I. 1997. Comparative evaluation of nitrogen release patterns from
756 controlled-release fertilizers by nitrogen leaching analysis. HortScience 32:669-
757 673.

758

759 Carroll, M.D. and J.F. Loneragan. 1969. Response of plant species to
760 concentrations of zinc in solution II. Rates of zinc absorption and their relation to
761 growth. Austral. J. Agric. Res. 20: 457-463.

762

763 Chen, J., R.C. Beeson, Jr., T.H. Yeager, R.H. Stamps and L.A. Felter. 2003.
764 Evaluation of captured rainwater and irrigation runoff for greenhouse foliage and
765 bedding plant production. HortScience 38:228-233.

766

767 Dancer, W.S., L.A. Peterson and G. Chesters. 1973. Ammonification and
768 nitrification of N as influenced by soil pH and previous N treatments. Soil. Sci.
769 Soc. Amer. Proc. 37:67-69.

770

771 Diamond, D. 1995. Phosphorus in soil extracts. QuikChem method 10-115-01-1-
772 A. Lachat Instruments, Milwaukee, Wis.

773

774 Dunham, C.W., and D.V. Tatnall. 1961. Mineral composition of leaves of three
775 holly species grown in nutrient sand cultures. Proc. Amer. Soc. Hort. Sci. 78:564-
776 571.

777

778 Edwards, D.G., and B.D. Horton. 1981. Influence of magnesium concentrations
779 in nutrient solution on growth, tissue concentration, and nutrient uptake of peach
780 seedlings. J. Amer. Soc. Hort. Sci. 106:401-405.

781

782 Engelsjord, M.E., O. Fostad, and B.R. Singh. 1997. Effects of temperature on
783 nutrient release from slow-release fertilisers. 1. Commercial and experimental
784 products. Nutrient Cycling in Agroecosystems 46:179-187.

785

786 Flint, H.L. 1962. Effects of different soil levels and methods of application of
787 phosphorus on growth of selected woody ornamental species in containers. Proc.
788 Amer. Soc. Hort. Sci. 81:552-555.

789

790 Foster, W.J., R.D. Wright, M.M. Alley, and T.H. Yeager. 1983. Ammonium
791 adsorption on a pine-bark growing medium. J. Amer. Soc. Hort. Sci. 108:548-
792 551.

793

794 Franson, M.A. 1985. Method 303A, p. 157-160. In: M.A. Farnson (ed.). Standard
795 methods for the examination of water and wastewater, 16th edition. Amer. Public
796 Health Assn. – Amer. Water Works Assn. - Water Pollution Control Federation,
797 Washington, D.C.

798

799 Frost, M.D., J.C. Cole and J.M. Dole. 2003. Fertilizer source affects iron,
800 manganese, and zinc leaching, nutrient distribution, and geranium growth. J.
801 Plant Nutr. 26:315-329.

802

803 Gandeza, A.T., S. Shoji and I. Yamada. 1991. Simulation of crop response to
804 polyolefin-coated urea: I. Field dissolution. Soil Sci. Soc. Amer. J. 55:1462-1467.

805

806 Gartner, J.B., M.M. Meyer, Jr., and D.C. Saupe. 1971. Hardwood bark as a
807 growing media for container-grown ornamentals. Forest Products J. 21:25-29.

808

809 Goertz, H.M. 1993. Controlled release technology. p. 251-274. In: Kirk-Othmer
810 (ed.). Encyclopedia of chemical technology. Vol. 7, 4th ed. Wiley & Sons, New
811 York, NY.

812

813 Handreck, K.A. 1989. Assessment of iron availability in soilless potting media.
814 Commun. Soil Sci. Plant Anal. 20:1297-1320.

815

816 Handreck, K.A. 1996. Phosphorus immobilization in wood waste-based potting
817 media. *Commun. Soil Sci. Plant Anal.* 27:2295-2314.
818
819 Handreck, K. 1997. Controlled release fertilizers. Measuring nutrient release
820 rates. *Austral. Hort.* 95(10):51-53.
821
822 Handreck, K. and K. Bunker. 1996. Fertilisers and hot weather. *Austral. Hort.*
823 94(8):83-85.
824
825 Havis, J.R. and J.H. Baker. 1985a. Phosphorus availability in peat-sand media
826 fertilized with several phosphorus sources. *J. Environ. Hort.* 3:153-155.
827
828 Havis, J.R. and J.H. Baker. 1985b. Phosphorus requirement of *Rhododendron*
829 'Victor' and *Cotoneaster adpressa praecox* grown in a perlite-peat medium. *J.*
830 *Environ. Hort.* 3:63-64.
831
832 Holcomb, E.J. 1981. Potassium release from selected slow-release fertilizers.
833 *Commun. Soil Sci. Plant Anal.* 12:1303-1310.
834
835 Huett, D.O. 1997a. Fertiliser use efficiency by containerized nursery plants. 1.
836 Plant growth and nutrient uptake. *Austral. J. Agr. Res.* 48:251-258.
837

838 Huett, D.O. 1997b. Fertilizer use efficiency by containerized nursery plants. 2.
839 Nutrient leaching. *Austral. J. Agr. Res.* 48:259-265.
840
841 Huett, D.O. and S.C. Morris. 1999. Fertiliser use efficiency by containerized
842 nursery plants. 3. Effect of heavy leaching and damaged fertilizer prills on plant
843 growth, nutrient uptake, and nutrient loss. *Austral. J. Agric. Res.* 50:217-222.
844
845 Huett, D.O. and B.J. Gogel. 2000. Longevities and nitrogen, phosphorus, and
846 potassium release patterns of polymer-coated controlled-release fertilizers at
847 30°C and 40°C. *Commun. Soil Sci. Plant Anal.* 31:959-973.
848
849 Husby, C.E., A.X. Niemiera, J.R. Harris, and R.D. Wright. 2003. Influence of
850 diurnal temperature on nutrient release patterns of three polymer-coated
851 fertilizers. *HortScience* 38:387-389.
852
853 Ingram, D.L. 1981. Characterization of temperature fluctuations and woody plant
854 growth in white poly bags and conventional black containers. *HortScience*
855 16:762-763.
856
857 Ivy, R.L., T.E. Bilderback and S.L. Warren. 2002. Date of potting and fertilization
858 affects plant growth, mineral nutrient content, and substrate electrical
859 conductivity. *J. Environ. Hort.* 20:104-109.
860

861 Johnson, C.M. and A. Ulrich. 1959. Analytical methods for use in plant analysis.
862 Univ. Calif. Agr. Expt. Sta. Res. Bul. 766.
863

864 Kochba, M., S. Gambash, and Y. Avnimelech. 1990. Studies on slow release
865 fertilizers. 1. Effect of temperature, soil moisture and water vapor pressure. Soil
866 Sci. 149:339-343.
867

868 Lamont, G.P., R.J. Worrall and M.A. O'Connell. 1987. The effects of temperature
869 and time on the solubility of resin-coated controlled-release fertilizer under
870 laboratory and field conditions. Scientia Hort. 32:265-273.
871

872 Lunt, O.R. and J.J. Oertli. 1962. Controlled release of fertilizer minerals by
873 encapsulating membranes: II. Efficiency of recovery, influence of soil moisture,
874 mode of application, and other considerations related to use. Soil Sci. Soc. Amer.
875 Proc. 26:584-587.
876

877 Mankin, K.R. and R.P. Fynn. 1996. Nutrient uptake response of New Guinea
878 impatiens to light, temperature, and nutrient solution concentration. J. Amer. Soc.
879 Hort. Sci. 121:826-830.
880

881 Niemiera, A.X. and R.D. Wright. 1986. The influence of nitrification on the
882 medium solution and growth of holly, azalea, and juniper in a pine bark medium.
883 J. Amer. Soc. Hort. Sci. 11:708-712.

884

885 Oertli, J.J. and O.R. Lunt. 1962. Controlled release of fertilizer minerals by
886 encapsulating membranes: I. Factors influencing the rate of release. *Soil Sci. Soc.
887 Amer. Proc.* 26:579-583.

888

889 Patel, A.J. and G.C. Sharma. 1977. Nitrogen release characteristics of
890 controlled-release fertilizers during a four month soil incubation. *J. Amer. Soc.
891 Hort. Sci.* 102:364-367.

892

893 Prasad, M. and M.J. Woods. 1971. Release characteristics of nitrogen fertilizer in
894 peat and sand. *J. Agr. Food Chem.* 19:96-98.

895

896 Rathier, T.M. and C.R. Frink. 1989. Nitrate in runoff water from container grown
897 juniper and Alberta spruce under different irrigation and N fertilization regimes. *J.
898 Environ. Hort.* 7:32-35.

899

900 Sweeney, R.A. 1989. Generic combustion method for determination of crude
901 protein in feeds: Collaborative study. *J. Assn. Offic. Anal. Chem.* 72:770-774.

902

903 Thomas, S. and F.B. Perry, Jr. 1980. Ammonium nitrogen accumulation and
904 leaching from an all pine bark medium. *HortScience* 15:824-825.

905

906 Schnitzer, M. and S.I.M. Skinner. 1966. Organo-metallic interactions in soils: 5.
907 Stability constants of Cu^{++} , Fe^{++} , and Zn^{++} -fulvic acid complexes. Soil Sci.
908 102:361-365.
909

910 Starr, K.D. and R.D. Wright. 1984. Calcium and magnesium requirements of *Ilex*
911 *crenata* 'Helleri'. J. Amer. Soc. Hort. Sci. 109:857-860.
912

913 Thurston, R.V., R.C. Russo, and K. Emerson. 1974. Aqueous ammonia
914 equilibrium calculations. Fisheries Bioassay Lab. Tech. Rpt. No. 74-1, Mont.
915 State Univ., Bozeman, Mont.
916

917 Tinus, R.W. and S.E. McDonald. 1979. How to grow tree seedlings in containers
918 in greenhouses. USDA For. Serv. Gen. Tech. Rep. RM-60.
919

920 U.S. Environmental Protection Agency. 1976. Quality Criteria for Water. EPA
921 Document No. 440/9-76-023. United States Environmental Protection Agency,
922 Washington D.C.
923

924 U.S. Environmental Protection Agency. 1994. Water quality standards handbook.
925 2nd ed. United States Environmental Protection Agency, Washington, D.C.
926

- 927 Walden, R.F. and R.D. Wright. 1995. Interactions of high temperature and
928 exposure time influence nitrification in a pine bark medium. HortScience 30:1026-
929 1028.
- 930
- 931 Wright, R.D. 1984. The pour-through method: a quick and easy way to determine
932 a medium's nutrient availability. Amer. Nurseryman 160(3):109-111.
- 933
- 934 Yeager, T.H. and R.D. Wright. 1982. Phosphorus requirement of *Ilex crenata*
935 Thumb. cv. Helleri grown in pine bark medium. J. Amer. Soc. Hort. Sci. 107:558-
936 562.